

Decomposition kinetics of ammonia in gaseous stream by a nanoscale copper-cerium bimetallic catalyst

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Abstract

This study performance is to examine the kinetics over nanoscale copper-cerium bimetallic catalyst under selective catalytic oxidation (SCO) of ammonia to N_2 in a tubular fixed-bed reactor (TFBR) at temperatures from 150 to 400 °C in the presence of oxygen. The nanoscale copper-cerium bimetallic catalyst was prepared by co-precipitation with $Cu(NO_3)_2$ and $Ce(NO_3)_3$ at molar ratio of 6:4. Experimental results showed that the catalyst with transmission electron microscopy (TEM) revealed that copper and cerium are well dispersed and catalyst in the form of nanometer-sized particles. Moreover, the kinetic behavior of NH_3 oxidation with catalysis can be accounted by using the rate expression of the Langmuir–Hinshelwood type kinetic model. Kinetic parameters are also developed on the basis of the differential reactor data. Also, experimental results are compared with those of the model predicted.

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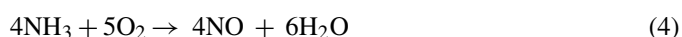
Keywords: Selective catalytic oxidation (SCO); Tubular fixed-bed reactor (TFBR); Ammonia; Nanoscale copper-cerium bimetallic catalyst; Kinetic

1. Introduction

Ammonia (NH_3) is widely found in industrial processes and it is a valuable chemical product for using with a variety of purposes and its can be eliminated in several ways, including ammonium nitrate production industry, livestock feedlots, urea manufacture plants, nitrogen fertilizer application industry, fossil fuel combustion or petroleum refineries, and refrigeration industry, which are either toxic inorganic gaseous and its pungent malodorous component under ambient conditions or have the potential harmful effects on the public [1–4]. It is also one of the control lists of the EPA, Taiwan. Moreover, conventional biological, physical, and chemical treatment processes, such as biofilters [5,6], stripping [7], water scrubbing [8], post-combustion control technologies [9], microwave-plasma discharge [10], electrochemical oxidation [11], and activated carbon fibers on soot adsorption [12,13], only make a phase transformation and may yield a contaminated sludge and/or an adsorbent, both needing for further disposal, and their maintenance or operation costs are high by physical and/or chemical

methods. Therefore, the removal, control, and prevention of ammonia emission from air and waste streams are important. This discharge presents a challenging task because environmental laws and regulations regarding safe discharge are becoming increasingly stringent is a major concern.

More recently, catalytic oxidation has been known to enhance the possibilities of the advanced oxidation processes technology through the use of dedicated catalysts, which can potentially promote oxidation in shorter reaction times under milder operating conditions. The selective catalytic oxidation (SCO) for ammonia-containing stream to molecular nitrogen and water is one way to solve ammonia pollution problems [14–19]. There are some reports about catalytic oxidation of ammonia that can occur, and these are as follows:



There are not many types of catalysts that have been used for oxidation of ammonia in gaseous phase. For example, Amblard et al. [3] have demonstrated excellent selective conversion of

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ammonia to nitrogen (>90%) by γ -Al₂O₃-supported Ni in selective catalytic oxidation processes. Furthermore, Wang et al. [15], who developed Ni-based catalysts for fuel gas oxidation from biomass gasification, found fresh Ni-based catalysts to be more active at lower temperature for the decomposition process of ammonia, and the partial pressure of hydrogen in the flue gas was a key factor to model ammonia oxidation. Liang [18] studied ammonia oxidation in a fixed-bed microreactor in the temperature range of 600–750 °C at GHSV = 1800–3600 h⁻¹. He observed that the conversion of ammonia reached 98.7 and 99.8% on nitrated MoNx/ α -Al₂O₃ and NiMoNy/ α -Al₂O₃ catalysts, respectively. Olofsson et al. [19] have demonstrated that excellent catalytic conversion of ammonia for nitrogen formation and by γ -Al₂O₃-supported Pt/CuO in selective catalytic oxidation processes. Among these, Schmidt-Szałowski [14] also published a paper covering a hypothetical model to explain the effect, the activity, and the selectivity in ammonia oxidation of the cobalt oxide catalyst's macrostructure on its properties.

As concerns reaction kinetic model, Lou [20] used a catalyst composed of Pt, Ni and Cr alloy of foam type to study the kinetics of catalytic incineration of butanone and toluene. He found that the Mars and van Krevelen model was suitable to describe the catalytic incineration of those volatile organic compounds (VOCs). Lou [21] used a Pt/Al₂O₃ alloy catalyst to study the kinetics of catalytic incineration of trichloromethane. He adopted power-rate law kinetics and found that the reaction was first-order in trichloromethane concentration and the activation energy was 16.2 kcal/mol. Lou [22] also used a 0.05% Pt/Ni/Cr alloy catalyst to study the kinetics of catalytic incineration of trichloromethane. He found that the Mars and van Krevelen model was suitable to describe the catalytic incineration of these VOCs. Gangwal [23] used a 0.1% Pt, 3% Ni/ γ -Al₂O₃ catalyst to study the kinetics of deep catalytic oxidation with *n*-hexane and benzene. They found that the Mars and van Krevelen model was favorable to explain the catalytic combustion of a binary mixture at temperature ranging from 160 to 360 °C. Nitrogen compounds have been shown the similar manner, which were reversible inhibitors to the catalyst [24]. However, the kinetic studies of catalytic oxidation of NH₃ on metal composite catalysts have not been thoroughly investigated.

In this study, we investigate the nature of the adsorbed species formed on the catalyst surface using an interpretation of the kinetic data. Various kinetic models, including the power-rate law, the Mars and van Krevelen model, and the Langmuir–Hinshelwood model were evaluated in driving the rate expression for NH₃ oxidation. Hence, we sought to study the activity of the nanoscale copper-cerium bimetallic catalyst on oxidation of ammonia at various parameters and the kinetic behavior of ammonia removal in the effluent stream. Our results can provide valuable information for designing and treating an ammonia-related system.

2. Materials and methods

In the preparation of the nanoscale copper-cerium bimetallic catalysts used in this study were prepared by co-precipitation of copper(II) nitrate (GR grade, Merck, Darmstadt, Germany), and

cerium(III) nitrate (GR grade, Merck, Darmstadt, Germany) at molar ratio of 6:4. They were then calcined at 500 °C in an air stream for 4 h. The resulting powder was made into tablets using acetic acid as a binder. The tablets were later reheated at 300 °C to burn the binder out of the nanoscale copper-cerium bimetallic tablets. They were then crushed and sieved into various particle sizes ranging between 0.25 and 0.15 mm for later use.

The experiments were carried out on a tubular fixed-bed flow quartz reactor (TFBR). Four flowing gases, namely NH₃, prepared the feeding mixture and the diluent's gas, helium, at the inlet of the reactor, each of them was independently controlled with a mass flow regulator. High purity dry air was used as the carrier gas and controlled with a mass flow meter (Sierra instrument Inc., USA) in the range of 8–13 L/min. The weight of catalysts is 1 g (empty bed volume approximately at 1.2 cm³). An inert material of γ -Al₂O₃ spheres (a hydrophilic inert material) was used to increase the interfacial area in gas phase for better mass transfer of ammonia from air. Hence, the limitations of external mass transfer and inter-particle diffusion were negligible in this reactor. The reaction tube (length 300 mm and inner diameter 28 mm) was placed inside a split tube furnace; and the tube containing the catalyst was also placed inside a split tube furnace. Two thermocouples of type K (Omega, USA) and a diameter of 0.5 mm were mounted and equally spaced along the catalyst bed. The thermocouples were also connected to a PID controller (FP21, Shimadzu Co Ltd, Japan) to maintain a uniform temperature in the tube within $\pm 0.5\%$. The feed gas (GHSV, 92,000 ml/h-g) was controlled at the concentration of 1000 ppm NH₃ under a GHSV of 92,000 ml/h-g and a concentration of O₂ of 4%. The catalyst's deactivation was not apparent during the tests. The reaction rate (*r*) of NH₃ is defined as follows:

$$r = \frac{F}{V} \cdot (C_{in} - C_{out}) = \left(\frac{F}{V}\right) \cdot C_{in} \cdot X \quad (5)$$

in which *r* is the reaction rate of NH₃ (mol ml⁻¹ s⁻¹), *F/V* is the space velocity per second, *C*_{in} and *C*_{out} are the inlet and outlet concentration (mol ml⁻¹) of NH₃, *X* is the conversion of NH₃ (%).

To study the oxidation kinetics expressions involving surface interactions and competitive adsorption of NH₃ over nanoscale copper-cerium bimetallic catalyst with highly active, three kinetic models: power-rate law model, Mars and van Krevelen model, and Langmuir–Hinshelwood model were examined [25–27]. These approaches will be adopted in the next section in this paper. The following overall rate equations were found using the Langmuir–Hinshelwood model (Eqs (6) and (7)) and the Mars and van Krevelen model (Eqs (8)–(11)) to describe the oxidation of ammonia, respectively.

$$-r_i = \frac{kK_i C_i K_{oi} C_{oi}}{(1 + K_i C_i + K_{oi} C_{oi})^2} \quad (6)$$

Eq (6) can be simplified to a linear function:

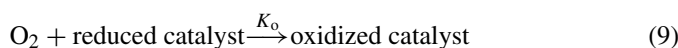
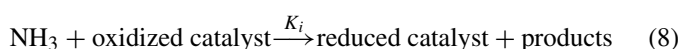
$$\sqrt{\frac{C_i}{-r_i}} = \frac{1}{\sqrt{K'}} + \frac{K''}{\sqrt{K'}} C_i \quad (7)$$

where

$$K_1' = \frac{kK_i C_i K_{oi}}{(1 + K_i C_i)^2}, \quad K_1'' = \frac{K_{oi}}{1 + K_i C_i},$$

$$K_2' = \frac{kK_i C_i \sqrt{K_{oi}}}{(1 + K_i C_i)^2}, \quad K_2'' = \frac{\sqrt{K_{oi}}}{1 + K_i C_i}$$

Moreover, the Langmuir–Hinshelwood model represents reactions in which the NH_3 can only adsorb chemically onto specific sites of the nanoscale copper-cerium bimetallic catalyst, each site can only adsorb unimolecular or atom and the adsorption is limited to a monolayer. Furthermore, Mars and van Krevelen reaction is adopted and shows the following sequences:



In this mechanism, the above steps are assumed to be first-order in the respective gaseous species and then Mars and van Krevelen can derive the following rate expression from a classical kinetic treatment

$$-r_i = \frac{K_i C_i K_{oi} C_{oi}}{K_i C_i + \alpha_i K_{oi} C_{oi}} \quad (10)$$

Eq (10) can be simplified to a linear function:

$$\frac{1}{-r_i} = \frac{\alpha_i}{K_{oi} C_{oi}} + \frac{1}{K_i C_i} \quad (11)$$

where k , K_{oi} , K_i are constants. K_i is the surface reduction rate constant (s^{-1}), K_{oi} is the surface reoxidation rate constant (s^{-1}), C_i is the oxygen concentration, and C_{oi} is the NH_3 inlet concentration. When $K_{oi} C_{oi} \gg K_i C_i$ in Eq (6) and $K_i C_i \gg \alpha_i K_{oi} C_{oi}$ in Eq (10) two rate equations can be expressed as the power-law form; they will be discussed below. Accordingly, the Mars and van Krevelen mechanism reaction between adsorbed oxygen and an adsorbed reactant of catalytic oxidation over metal oxides is also a redox mechanism involving gas-phase and lattice oxygen as well as gas-phase oxygen.

Samples before and after reaction were injected automatically with a sampling valve into a gas chromatograph (Shimadzu GC-14A) equipped with a thermal conductivity detector. A stainless-steel column (Porapak Q 80/100 mesh) was used for separation and analysis of the concentrations of N_2O isothermally at 100°C . NH_3 was injected automatically before and after the reaction with a sampling valve into a gas chromatograph (Shimadzu GC-14A) equipped with a TCD (Shimadzu, Kyoto, Japan). The signal areas were measured electronically with a data integrator (CR-6A, Shimadzu). The concentrations of NO , NO_2 and O_2 in the gas samples were measured by a portable flue gas analyzer (IMR3000, IMR, Germany), which

was linked to the designed location for continuous monitoring during combustion. Gas samples with known concentrations of NH_3 were used for the calibration.

Transmission electron microscopy (TEM) (Philips CM-200 Twin, Netherlands) elucidated the morphology of the catalysts and yielded information on the distribution of copper-cerium species on the catalyst surface. UV–vis absorption spectra were performed with a photo spectrophotometer (MCPD-3000, OTSUKA Electronics, Osaka, Japan) for the solid sample. The measured reflectance is relative to a BaSO_4 standard.

3. Results and discussions

Table 1 lists the test conditions: the concentration of the single compound, the space velocity, and the operating temperature. The procedures consisted of experiments with a differential reaction. Differential reaction experiments were performed to investigate the various kinetic models and it was evaluated to derive the rate expression for the NH_3 oxidation.

Fig. 1 presents the surface morphological changes of nanoscale copper-cerium bimetallic catalyst elucidating using TEM to provide information on the fresh/aged catalyst surface structure. Fig. 1a indicates that nanoscale copper-cerium bimetallic exhibits more aggregation and crystalline than observed in Fig. 1b. These crystal phases may explain the high activity of the catalysts. Fig. 1b indicates that disaggregated and dispersed phases were formed when the surface of the catalyst was aged or when poisoning occurred because of plugging, implying that the porosity of the particles had changed. These crystal phases may be responsible for the high activity of the catalysts. These results also confirm that the dispersion phenomena of the nanoscale copper-cerium bimetallic catalyst increase the efficiency of removal of NH_3 .

To gain further information on the state of copper-cerium species in these catalysts, the UV–vis spectroscopic studies are carried out and the spectra are shown in Fig. 2. Generally, two bands were observed for these copper-cerium catalysts at 330 and 600–800 nm. The band at 330 nm is attributed to the $\text{Cu}^{2+}\text{-O}^{2-}$ electronic transition species. The band at 600–800 nm is assigned to the octahedrally coordinated Cu^{2+} in the literature [28]. The formation of these species is in good with Carvalho [28], who indicated the reductive effect of NH_3 on nanoscale copper-cerium catalyst. The band at 600–900 nm is also assigned to $d\text{-}d$ transitions of Cu^{2+} situated in an octahedral environment with an O_h symmetry [29]. In addition, bulk CeO_2 also shows strong absorption in this region. We thus prefer to assign the band at 340–350 nm to the bulk-like CeO_2 species. CeO_2 in a copper-cerium bimetallic catalyst may be assumed to promote the generation of the active phase of CuO under the conditions of NH_3 oxidation.

Table 1
Experimental conditions of test gas

Reaction type	Space velocity (1/h)	Residence period (s)	NH_3 concentration (ppm)	Temperature ($^\circ\text{C}$)	Oxygen content (%)	Temperature rises (DT)
differential	92,000	0.04	1000	150–400	4	<1

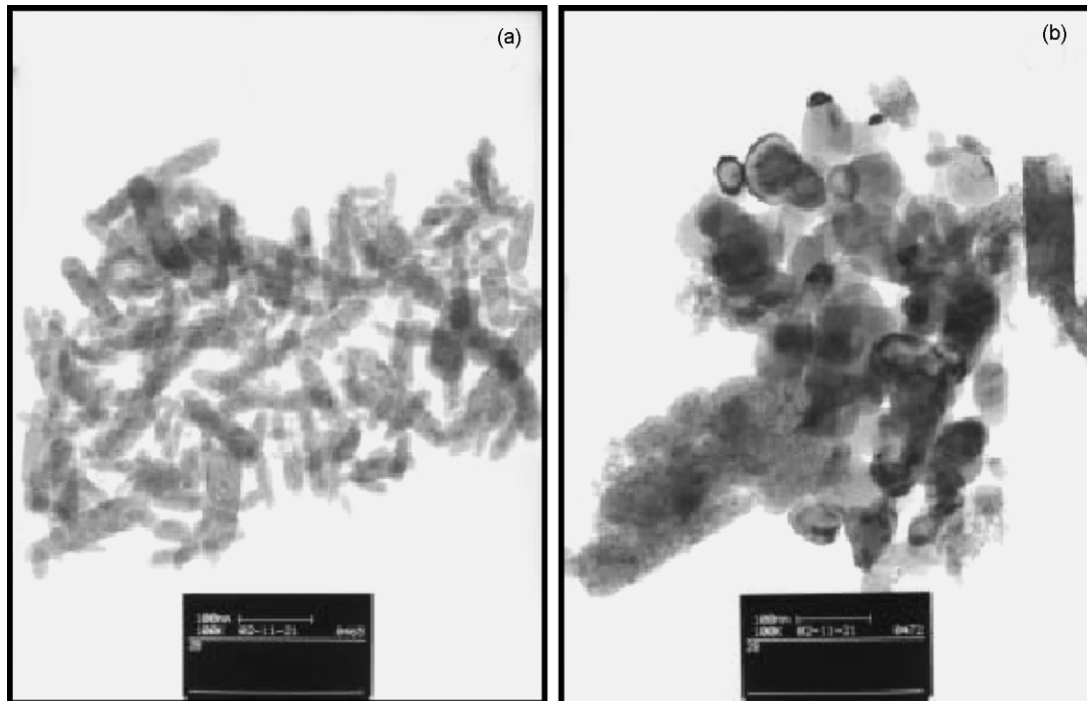


Fig. 1. TEM photograph of (a) fresh and (b) after activity test nanoscale copper-cerium bimetallic catalyst. Test conditions: 1000 ppm NH_3 in He, $\text{O}_2 = 4\%$, GHSV = 92,000 ml/h-g.

The measured rates of oxidation of NH_3 in air stream were correlated first using the empirical power rate law kinetics. Then the dependence of the rate of NH_3 oxidation on the concentration of NH_3 and oxygen was tested using the following power-rate law equation:

$$-r_i = kC_i^n C_{\text{O}_2}^m \quad (12)$$

in which k is the reaction constant, and m and n are reaction orders. Since $C_{\text{O}_2} > C_i$ in most systems of catalyst incineration, the term can be assumed as a constant. Table 2 shows the reaction rates of catalytic incineration of NH_3 for the fixed O_2

concentration, which were calculated from Eq (5). Therefore, Fig. 3 plots $-\ln(-r_i)$ versus $-\ln C_i$ for NH_3 oxidation at several temperatures. The reaction order of C_i can be obtained from the slope of each line in Fig. 4. The linearity of plots results reveals a hypothesis of approximately first-order kinetic with changing from 0.55 to 0.94 in the NH_3 oxidation. In addition, the reaction constant k can be obtained from the intercept of each line in Fig. 3. Next, plotting the $\ln K$ versus T^{-1} , as shown in Fig. 4, attains a straight line. Via its intercept and its slope, the values of the pre-exponential factor [$A = 2.02 \times 10^8 \text{ s}^{-1}$] and the

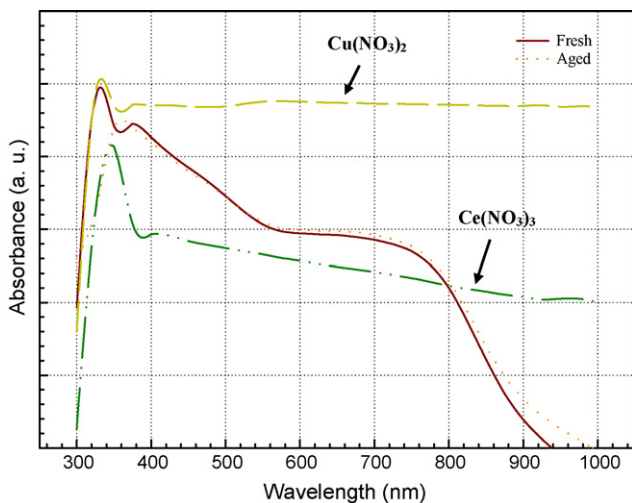


Fig. 2. UV-vis absorption spectra of (a) fresh and (b) after activity test nanoscale copper-cerium bimetallic catalyst. Test conditions: 1000 ppm NH_3 in He, $\text{O}_2 = 4\%$, GHSV = 92,000 ml/h-g.

Table 2

The reaction rates of catalytic incineration of ammonia of fixed O_2 concentration at 4%

Inlet temperature (°C)	Inlet NH_3 concentration (10^{-8} mol/cm^3)	Reaction rate ($10^{-7} \text{ mol/cm}^3/\text{s}$)
150	2.04	1.17
	2.45	1.40
	3.27	1.86
	4.08	2.31
200	2.04	1.41
	2.45	1.64
	3.27	2.15
	4.08	2.64
250	2.04	2.25
	2.45	2.45
	3.27	2.88
	4.08	3.31
300	2.04	2.69
	2.45	3.03
	3.27	3.55
	4.08	4.07

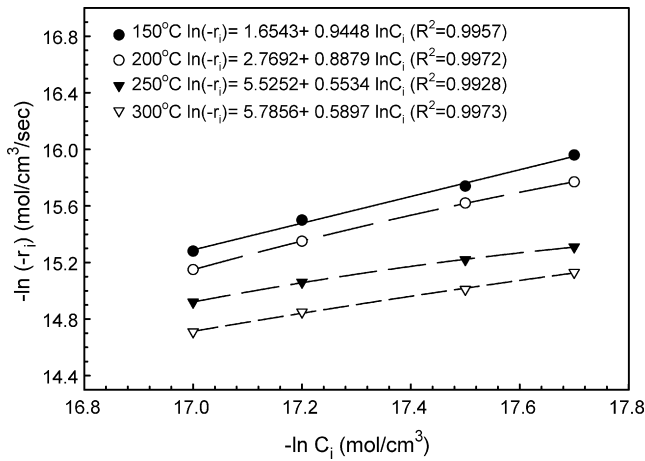


Fig. 3. Linear fitting of NH₃ kinetic data for the power-rate law.

apparent activation energy of NH₃ oxidation ($E_a = 62.1$ kJ/mol) were calculated from Arrhenius relationships. It is suggested that the power-rate law model was not feasible for the oxidation of ammonia over the nanoscale copper-cerium bimetallic catalyst since the R^2 value ($R^2 = 0.9285$) for the rate constants fitting is relatively low.

The dependence of the rate of NH₃ oxidation on the concentration of NH₃ and oxygen was tested using the following Mars and van Krevelen classical kinetic model of oxidation reaction with competition for surface adsorbed oxygen was assumed. In this study, Fig. 5 plots r_i^{-1} versus C_i^{-1} for NH₃ oxidation at several temperatures. The reaction order of C_i^{-1} can be obtained in this manner from the slope of each line in Fig. 5. The values of K_i and K_{oi} can be calculated from Eq (11) and the linearity of plots results reveals a reaction with changing from 0.50 to 1.65 in the NH₃ oxidation. In addition, the reaction constant k can be obtained from the intercept of each line in Fig. 5. Next, plotting the $\ln K$ versus T^{-1} , as shown in Fig. 6, attains a straight line. Via its intercept and its slope, the values of the pre-exponential factor [$A_i = 24.0$ s⁻¹; $A_{oi} = 2 \times 10^2$ s⁻¹] and the apparent activation energy of NH₃ oxidation, E_i , 12.2 kJ/mol and the apparent

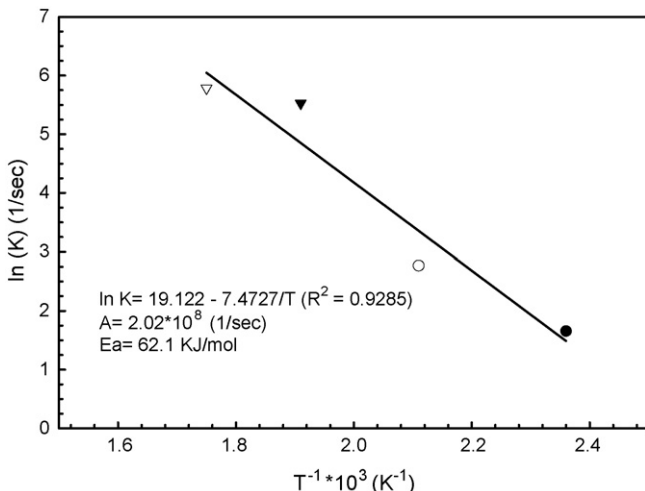


Fig. 4. Arrhenius equation fitting of NH₃ kinetic data for the power-rate law.

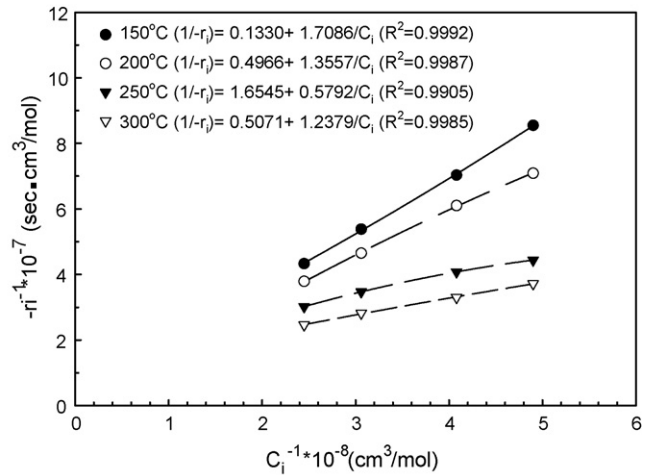


Fig. 5. Linear fitting of NH₃ kinetic data for the Mars and van Krevelen model.

activation energy of NH₃ oxidation, E_{oi} , 24.1 kJ/mol were calculated from Arrhenius relationships, respectively. If the ammonia react with the gas phase or adsorbed on different catalyst sites, a Mars and van Krevelen type kinetic model will be observed, and mutual inhibition is through competition for adsorbed oxygen. Consequently, it is suggested that the Mars and van Krevelen model was not feasible for the oxidation of ammonia over the nanoscale copper-cerium bimetallic catalysts since the R^2 value ($R^2 = 0.8213$ and 0.9151 , respectively) for the rate constants fitting is relatively low.

The dependence of the rate of NH₃ oxidation on the concentration of NH₃ and oxygen was tested using the following Langmuir–Hinshelwood model with reaction between oxygen and ammonia adsorbed over the surface sites. In this study, Fig. 7 plots $(C_i/-r_i)^{1/2}$ versus C_i for NH₃ oxidation at several temperatures. The R^2 value for various lines shows fairly good agreements. Table 3 shows the reaction rates of oxidation of NH₃ for fixed NH₃ concentration. In addition, the reaction constant K'_1 and K''_2 can be obtained from Eq (7) and the results are also given in Table 4. Fig. 8 plots $(C_{oi}/-r_{oi})^{1/2}$ versus C_{oi} for

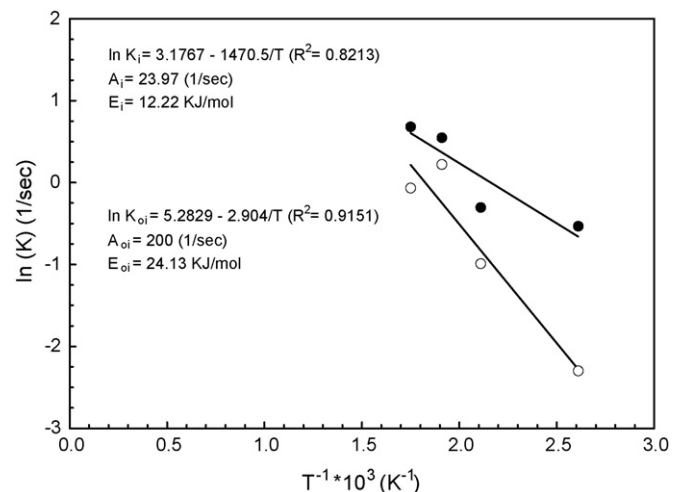


Fig. 6. Arrhenius equation fitting of NH₃ kinetic data for the Mars and van Krevelen model.

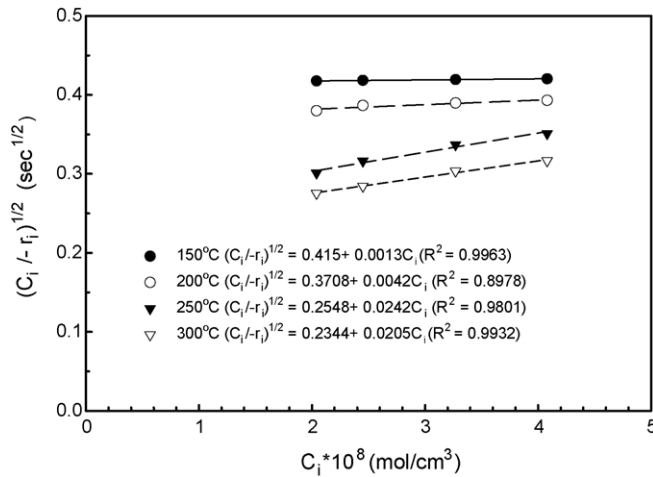


Fig. 7. Calculating K'_1 and K''_2 by linear fitting of NH_3 kinetic data for the Langmuir–Hinshelwood model.

Table 3

The reaction rates of catalytic incineration of ammonia fixed NH_3 concentration at 1000 ppm

Inlet temperature ($^{\circ}\text{C}$)	Inlet O_2 concentration (10^{-6} mol/cm 3 /s)	Reaction rate (10^{-8} mol/cm 3 /s)
150	1.63	5.55
	4.08	5.68
	6.94	5.78
	8.16	5.79
200	1.63	5.58
	4.08	5.67
	6.94	5.80
	8.16	5.87
250	1.63	6.01
	4.08	6.55
	6.94	6.71
	8.16	7.00
300	1.63	6.13
	4.08	6.59
	6.94	6.79
	8.16	7.27

Table 4

The composite rate constants of the Langmuir–Hinshelwood model for the reduction of ammonia

	Inlet temperature ($^{\circ}\text{C}$)			
	150	200	250	300
(1) Fixed O_2 concentration				
K'_1	5.8	7.9	15.4	18.2
$K''_2 \times 10^{-2}$	0.3	1.1	9.5	8.8
(2) Fixed NH_3 concentration				
(a) Molecular oxygen adsorption				
K'_1	0.05	0.05	0.06	0.06
K''_1	0.23	0.23	0.21	0.20
(b) Atomic oxygen adsorption				
$K'_2 \times 10^{-5}$	9.34	10.6	9.91	9.79
K''_2	0.37	0.45	0.36	0.34

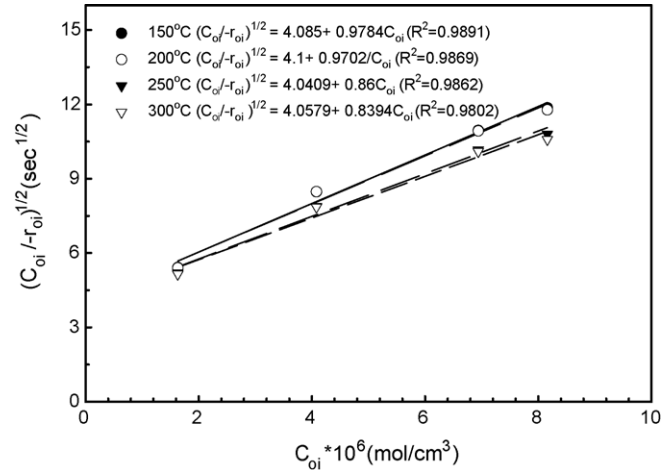


Fig. 8. Calculating K'_1 and K''_2 by linear fitting of NH_3 kinetic data for the Langmuir–Hinshelwood model involving molecular oxygen adsorption.

NH_3 oxidation at several temperatures and its involving molecular oxygen adsorption. The reaction order of C_{oi} can be obtained from the slope of each line in Fig. 8. The linearity of plots results reveals an approximately first-order reaction changing from 0.83 to 0.98 in the NH_3 oxidation. Via its intercept and its slope, the values of the pre-exponential factor [$A = 1.2 \times 10^{-7} \text{ s}^{-1}$] and the apparent activation energy of NH_3 oxidation ($E_a = 27.8 \text{ kJ/mol}$) were calculated from Arrhenius relationships in Fig. 9. Fig. 10 plots $((C_{oi})^{1/2}/r_{oi})^{1/2}$ versus $C_{oi}^{1/2}$ for NH_3 oxidation at several temperatures and its involving atomic oxygen adsorption. Accordingly, ammonia reacts from gas phase with adsorption oxygen atomic and molecular oxygen. If the surface reaction is much slower than the adsorption processes, then equilibrium adsorption will occur and a Langmuir–Hinshelwood type is observed [30]. The reaction order of C_{oi} can be obtained from the slope of each line in Fig. 10. Via its intercept and its slope, the values of the pre-exponential factor [$A = 0.36 \text{ s}^{-1}$] and the apparent activation energy of NH_3 oxidation ($E_a = 40.7 \text{ kJ/mol}$) were calculated from Arrhenius relationships in Fig. 11. The linearity of

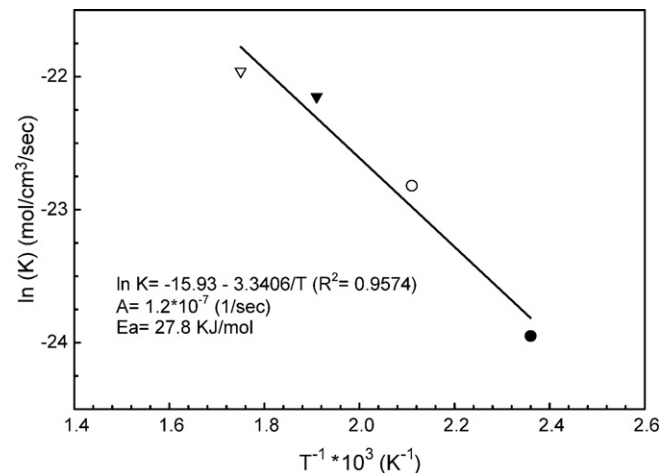


Fig. 9. Arrhenius equation fitting of NH_3 kinetic data for the Langmuir–Hinshelwood model involving molecular oxygen adsorption.

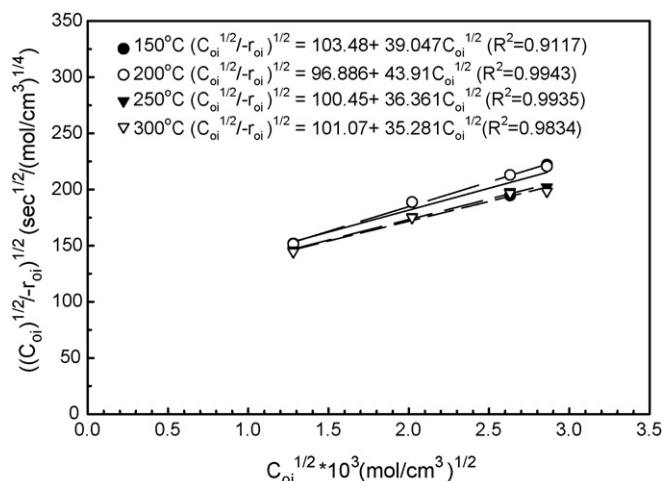


Fig. 10. Calculating K'_1 and K'_2 by linear fitting of NH_3 kinetic data for the Langmuir–Hinshelwood model involving atomic oxygen adsorption.

Table 5

The reaction rates constants and adsorption equilibrium constants of the Langmuir–Hinshelwood model for the reduction of ammonia

Inlet temperature (°C)	Molecular oxygen adsorption k ($10^9 \text{ mol/cm}^3/\text{s}$)	Atomic oxygen adsorption k ($10^8 \text{ mol/cm}^3/\text{s}$)
150	3.691	12.67
200	1.741	4.731
250	0.240	4.125
300	0.289	6.040

plots results reveals an approximately first-order reaction in the NH_3 oxidation. Hence, the linearity of plots results suggests the fitness of the Langmuir–Hinshelwood mechanisms for ammonia. Arrhenius relationships and plots for the corresponding k values for ammonia are shown in Figs. 9 and 11, respectively. Table 5 shows the reaction rates and adsorption equilibrium of oxidation of NH_3 for fixed NH_3 concentration. As a result, it is suggested that the Langmuir–Hinshelwood type kinetic model

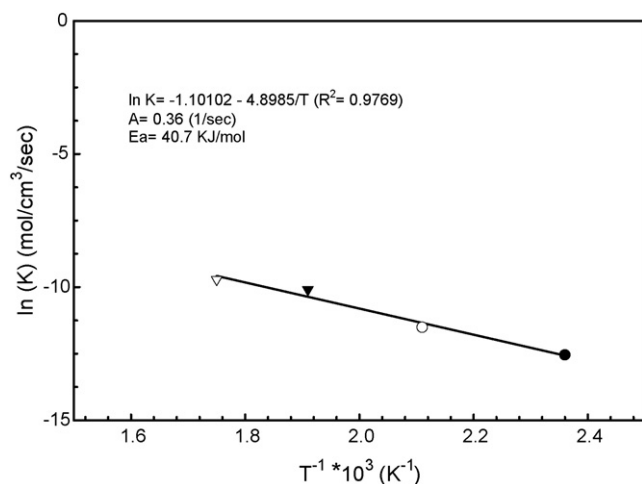


Fig. 11. Arrhenius equation fitting of NH_3 kinetic data for the Langmuir–Hinshelwood model involving atomic oxygen adsorption.

involving molecular oxygen and atomic oxygen adsorptive reactivity and competition adsorption surface sites remarkably were feasible for the oxidation of ammonia over the nanoscale copper-cerium bimetallic catalysts since the R^2 value ($R^2 = 0.9574$ and 0.9769 , respectively) for the rate constants fitting is relatively high. Thus, the proposed kinetic models proved to predict reaction mechanism and rates satisfactorily. The result is similar to that described in relevant literature [31–35].

Results showed that the kinetic behavior of NH_3 oxidation with catalysis could be explained by using the rate model of Langmuir–Hinshelwood. Regarding the rate expressions for the oxidation of NH_3 catalyzed by nanoscale copper-cerium bimetallic catalyst, the kinetic parameters such as K and E_a were obtained by the linear regression method, we can substituted into Arrhenius equation and an equation of the theoretical values of conversion at various reaction temperatures is

$$X = 1 - \exp[-2 \times 10^2 \cdot \exp\left(\frac{-24.13}{RT}\right) \cdot \tau] \quad (13)$$

in which X is the conversion, R is the gas constant, T is the temperature of the reactor and τ is the residence period. The relation between the residence period for actual entering conditions and at reactor conditions is given by

$$\tau = \tau_0 \cdot \left(\frac{T}{T_0}\right) \quad (14)$$

in which τ_0 and T_0 are the residence period and the absolute temperature of the stream entering the reactor, respectively. In most cases, our residence period was based on the feed for actual entering conditions.

Fig. 12 compares predicted and experimental conversions obtained by and isothermal differential fixed-bed reactor at temperatures ranging from 150 to 400 °C. This figure reveals that the model accurately predicts the NH_3 conversion and

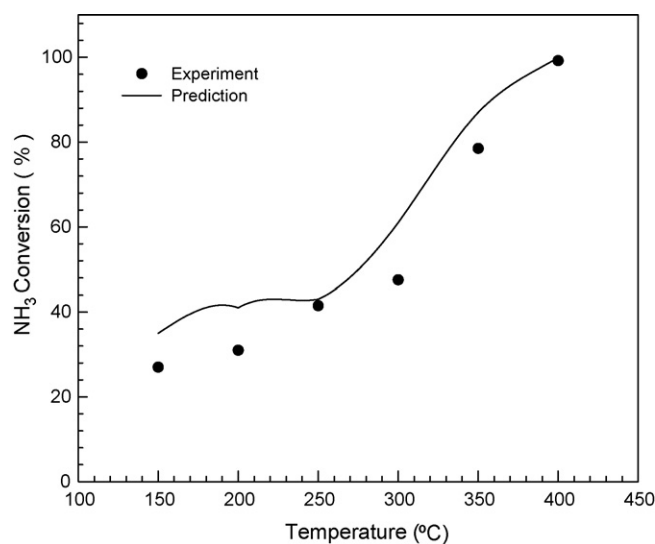


Fig. 12. Comparison of experimental data (symbols) and computed values (solid lines) from conversions of the rate equation at different temperatures of NH_3 by Langmuir–Hinshelwood type model. Test conditions: 1000 ppm NH_3 in He, $\text{O}_2 = 4\%$, GHSV = 92,000 ml/h-g.

the conversion of ammonia increased greatly from 38 to 99% (150–400 °C). The removal efficiency increases with increasing temperature and decreasing residence period. The deviation is enhanced, particularly in the region (250–400 °C) of a high concentration of NH₃ (1000 ppm). In this case, the temperature increase was measured indicating that the temperature effects of NH₃ in high concentration also slightly promoted the conversion of NH₃ oxidation; thus, conversions were slightly greater than the predicted values. This model is appropriate for oxygen presence conditions and high concentration of NH₃.

4. Conclusions

The catalytic oxidation of ammonia over the nanoscale copper-cerium bimetallic catalyst was found to lead to more effective ammonia removal containing air stream and selective catalytic process was found to be more effective at lower temperatures. A kinetic rate expression was developed to describe the data over the range of conditions investigated. Moreover, correlation equations in the form of Arrhenius's law are derived from experimental data to predict the destruction efficiencies of ammonia from the temperatures and residence period in a catalytic process. A Langmuir–Hinshelwood model incorporating a first-order reaction appears to adequately represent the NH₃ oxidation, which shows fairly good agreements with experimental data. The apparent activation energy is 27.8 kJ/mol. This kinetic model is appropriate for oxygen presence and high concentrations of NH₃. In this work, we use the catalytic oxidation showing a promise in the treatment of highly concentrated ammonia stream, thereby helping industrial plants meet the discharge regulations.

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